

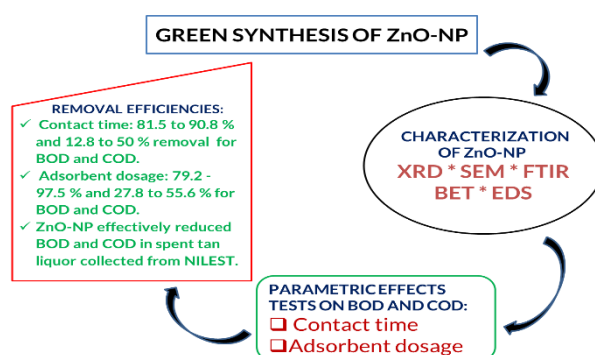
Synthesis, characterization and application of a bio-derived ZnO nanopowder from *spinacia oleracea* leaf extract for the removal of BOD and COD from tannery wastewater

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ABSTRACT

Biochemical Oxygen Demand (BOD) and Chemical Oxygen Demand (COD) as biological and chemical pollutants of wastewater are renowned environmental problems. A study of the performance of BOD and COD degradation via adsorption was undertaken using Bio-derived ZnO nanopowder (ZnO-NP), synthesized using leaf extract of *Spinacia oleracea*, and Zn (NO₃)₂ at 500 °C, following a simple and green approach. XRD, SEM, FTIR, EDS and BET analysis were used to characterize the nano-adsorbent. EDS spectrum recorded elemental weight compositions of 67.33% and 32.67% for Zn and O, while the FTIR absorption peaks revealed the presence of Zn-O-H and Zn-O. Surface area analysis revealed the mesoporous structure of the ZnO-NPs. The reduction efficiency of the ZnO-NPs was evaluated in the presence of raw tannery wastewater by application of treatment time and adsorbent dosage as parametric factors; results obtained were compared to environmental regulatory limits (WHO and NESREA). Contact times reported removal efficiencies of 81.5-90.8 % and 12.8-50 % for BOD and COD, while adsorbent dosage-influenced BOD and COD removal at an optimum contact time of 30 minutes was found to be 79.2 - 97.5 % and 27.8 to 55.6 % respectively. This study reveals that ZnO nanopowder is better applied as BOD reductants over COD for spent tan liquor.



HIGHLIGHTS

- EDS reveals the elemental weight of 67.33% and 32.67% for Zn and O, respectively.
- The ZnO-NP exhibited a stable photocurrent density.
- ZnO-NP recorded BOD removal efficiency of 81.5, 83.4, 85.8, 90.3 and 90.8 %.
- Removal efficiency of COD is low (12 - 50 %) but improves progressively with increasing contact time.

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1. Introduction

In recent times, the field of adsorption studies has witnessed additions of a variety of efficient, eco-friendly and cost-effective nano-materials, that finds application in the decontamination of industrial and domestic liquid wastes (Theron *et al.*, 2008; Gupta *et al.*, 2015; Shamsizadeh *et al.*, 2014; Kyzas and Matis *et al.*, 2015). Treatment of wastewater and drinking water can reduce contamination concerns (Ferroudj *et al.*, 2013); however, existing traditional methods

of treatment in practice are not efficient enough to completely remove emerging contaminants (Jiuhui, 2008; Ferroudj *et al.*, 2013; Bali *et al.*, 2003), thus, the emphasis on a more efficient, cheaper and powerful green technology for the treatment of municipal and industrial wastewaters (Jarvis, 2006; Ferroudj *et al.*, 2013). A variety of efficient, eco-friendly and cost-effective nano-materials with unique functionalities for potential decontamination of industrial effluents, surface water, groundwater and drinking water have been employed in wastewater treatment (Gupta *et al.*,

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2015; Theron *et al.*, 2008). From documented evidence, nano-technology has a proven ability to remediate wastewaters (Zare *et al.*, 2013; Sadegh *et al.*, 2014), because structurally, they are components with a dimensional size of at least below 100 nm (Amin *et al.*, 2014). BOD and COD levels in wastewaters are serious environmental concerns and could present detrimental effects especially when their presence exceeds environmental standards (Davi, 2009; Ramasami *et al.*, 2015)

Nanomaterials exist in a variety of forms, including nano-wires, nano-tubes, films, particles, quantum dots and colloids, all having excellent liquid treatment potential (Edelstein and Cammaratra, 1998; Lubick and Betts, 2008; Dutta and Maji, 2014). Nano-adsorbent can be produced using atoms of elements that are chemically active and having high adsorption capacity on the surface of the nano-material (Kyzas and Matis, 2015). Activated carbon, silica, clay materials, metal oxides and modified compounds in the form of composites are examples of nano-adsorbents employed for wastewater treatment (El-Saliby *et al.*, 2008).

The use of metal oxide-based nanoparticles as adsorbents is gaining recognition in environmental sciences. Researches on ZnO and CuO based nanoparticles for photocatalytic and biological activities have also been published (Saikia *et al.*, 2015; Tamuly *et al.*, 2014). Oladipo *et al.*, (2017), reported the synthesis of a bio-derived MgO non-powder and applied it in the reductive treatment of BOD and COD in tannery wastewater; with results showing a 93.5% and 96.9% removal. Because industrial activities vary, so also will the wastes they generate (both physical and chemical parameters), with pollution levels rising beyond prescribed levels. Untreated effluents when released into receiving water bodies (rivers, lakes, ponds, etc.), are a major contributor to water pollution (Rehman and Anjum, 2010). Spent Tan liquor from leather industries characteristically, in addition to metal contaminants contains high concentrations of organic particles and sediments, accounting for its high BOD (Biological Oxygen Demand) and COD (Chemical Oxygen Demand) levels (Ramasami *et al.*, 2015).

The contaminant parameter, BOD, is the quantity of oxygen consumed by organisms in breaking down liquid wastes or the amount of oxygen available to the organism to carry out its metabolic activities. COD on the other hand indicates the extent of organic matter contamination of a water system, and COD is always higher than BOD. COD is useful for understanding the overall organic load of a receiving water body (APHA, 1998; Sawyer *et al.*, 2000).

Tannery wastewater is a major source of environmental pollution, generating high organic pollutants (BOD and COD) at excessively high concentrations. BOD and COD are major organic pollution indicators of wastewater generated at various processing stages of the leather tanning process (e.g. soaking, unhairing/liming, lime fleshing, delimiting/bating, degreasing, pickling/tanning, sammying, retanning, dyeing and fatliquoring). Traditional materials

and treatment technologies (reverse osmosis (RO), oxidation, activated sludge, nanofiltration (NF), and activated carbon) are not adequate enough in the removal of huge amounts of organic micropollutants. The demand for organic micropollutants removal is the motivation for the research into economically feasible and environmentally stable water/wastewater treatment technology to meeting the environmental water quality standards. This study aims to synthesize and test the capacity of a biodegradable adsorbent (Zinc Oxide) in the remediative treatment of BOD and COD polluted spent tan liquor derived from the Nigerian Institute of Leather and Science Technology (NILEST) Zaria.

2. Materials and methods

2.1 Materials

Analytical grade chemicals were used without further purification in this study. Zinc nitrate hexahydrate (99.999 % trace metals basis) was procured from Thomas Scientific (Swedesboro, NJ), leaves of *Spinacia oleracea* from Samaru, Nigeria. Untreated tannery effluent was collected from the tannery at the Nigerian Institute of Leather and Science Technology, Zaria, Nigeria. The equipment used are listed below: FTIR (PerkinElmer 1750) was used to measure the infrared region of the electromagnetic radiation spectrum in order to identify chemical bonds and functional groups in the sample. BET analyzer (Quantachrome Nova 2000e) was used to measure the surface area of the adsorbent. Scanning electron microscope (JEM-ARM200F) was used to study the surface morphology and estimate elemental components of the adsorbent.

2.2 Collection and preparation of leave extract

Fresh leaves of *Spinacia oleracea* were obtained from the Samaru market in Zaria, Kaduna State. The leaves were washed thoroughly with tap water to remove sand and other solid particles and then rinsed severally with distilled water, chopped into tiny bits, air-dried for 12 days, grounded to powder using mortar and pestle. A-50 g of powdered leaves was weighed out and added to 400 mL of distilled water in a 1000 mL conical flask and stirred under heat at 80 °C for 30 minutes. The extract was mixed and centrifuged for 10 minutes and then filtered, the filtrate was kept in the refrigerator for use in the synthesis of ZnO-NP.

2.2.1 Green synthesis of ZnO nanopowders

An aliquot of 150 mL of the aqueous plant extract was mixed with 20 mL of $Zn(NO_3)_2 \cdot 6H_2O$ in a 400 mL conical flask. The mixture was vigorously stirred for 20 minutes at ambient temperature, forming milky precipitates and were collected by centrifugation. The precipitates were washed with ethanol, rinsed with distilled water, and dried at 90 °C for 3 hours. The dried precipitate powder was heated in an oven (Isco S4A00005-MICRA9S) at 500 °C for 2 hours to form ZnO nanopowder.

2.2.2 Characterization of ZnO nanopowders (ZnO-NPs)

The *spinacia oleracea* leaf was determined to be rich in chloride, potassium and carbonate ions after titration and elementary analyses. The ions present in the leaf (Na^+ , CO_3^{2-} , K^+) react with $\text{Zn}(\text{NO}_3)_2 \cdot 6\text{H}_2\text{O}$ to form $\text{Zn}(\text{OH})_2$ and the product was subjected to heat treatment at $500\text{ }^\circ\text{C}$ to form ZnO NPs.

2.3 Collection of tannery effluent

The tannery effluent was collected by spot (grab) sampling method as detailed elsewhere (UNIDO, 2016). Collected wastewater was filtered and stored in a cooling set, at freezing temperature (below $4\text{ }^\circ\text{C}$) till use.

2.3.1 Sorptive treatment experiment via batch and fixed-bed system

The tannery wastewater was treated according to the sorptive remediation method by Oladipo et al., (2017). The sorption experiment was carried out at ambient temperature under different experimental conditions (contact time and adsorbent dosage). Effect of contact time was achieved by contacting 50.0 mg of ZnO-NPs with 50 mL tannery wastewater of initial BOD of 91.8 mg/L and COD of 576 mg/L . The experiment was carried out in sealed conical flasks at 20 min interval between 20 to 100 minutes per experiment batch. Samples were drawn every 20 minutes , filtered and analyzed. Effect of adsorbent dosage was deduced at different nanopowder dosages (between 20 - 100 mg/L), at $\text{pH } 12$ (pH of the tan liquor). The above experimental procedures were carried out in triplicate from which the mean result was calculated. The percentage removal of COD and BOD concentrations were validated by using the equations:

$$R (\%) = \frac{(BOD_i \times V_i) - (BOD_f \times V_f)}{BOD_i \times V_i} \times 100 \quad (1)$$

$$R (\%) = \frac{(COD_i \times V_i) - (COD_f \times V_f)}{COD_i \times V_i} \times 100 \quad (2)$$

Where: BOD_i , COD_i , BOD_f and COD_f are the initial and final concentrations at a given time, respectively. V_i and V_f are the volumes of the untreated and treated wastewater solutions, respectively.

2.4 Determination of Organic Waste Contaminants (BOD and COD)

The BOD and COD of untreated and treated effluents were analyzed according to the standard methods of APHA (1998). In the BOD analysis, dilutions were prepared as appropriate, for the sample(s) to be tested. The diluted samples were transferred to corresponding glass stoppered BOD bottle(s), heated to $20\text{ }^\circ\text{C}$, then the DO of the sample was measured in mg/L using a DO meter and electrode. An appropriate quantity of nitrification inhibitor was added to the sample, and the bottle was again stoppered and incubated. The incubated samples were again tested for DO after the fifth day. The final DO reading was subtracted from the initial DO reading, and the result was the BOD

concentration (mg/L). The BOD_5 was determined using: $BOD_5 = DO_0 - DO_5$.

Before starting COD analysis, a series of known standards are prepared using potassium hydrogen phthalate (KHP) in the high range of 100 , 250 , 500 and 1000 mg/L . A COD reactor/heating ($150\text{ }^\circ\text{C}$) block and a colourimeter were turned on so that both instruments are allowed to stabilize. Seven pre-prepared high-range (20 - 1500 mg/L) vials were selected for the test and labelled as blank (1 vial), standard (4 vials) and wastewater (2 vials), and to each vial, 2 mL of deionized water, KHP standard and tannin liquor were added, respectively, mixed well and placed inside the reactor block for 2 h . After two hours, the vials were transferred to a cooling rack for about 15 minutes before colourimetric measurement. The colourimeter was set and calibrated as per the specific instructions for that unit (proper wavelength for blank and standards). Each vial was placed in the unit and the COD concentration was noted.

3. Results and discussion

The result of X-Ray Diffraction as presented in Fig. 1, the ZnO-NP was characterized by powder X-ray diffraction (equipped with $\text{Cu K}\alpha$ radiation (1.5406 \AA) and a graphite monochromator operated at 40 Kv and 30 mA in Bragg-Brentano geometry. The XRD spectra were collected in the 2θ range between 20° and 80° using a step-scan mode of 0.05° (2θ). Following synthesis, X-Ray diffractogram of the ZnO-NP exhibits high crystalline nature, with sharp and intense peaks. The intense diffraction peaks appeared at 31.9° , 34.8° , 36.6° , 48.3° , 56.9° , 63.9° , 67.9° , 69.2° , which corresponded to the (100), (002), (101), (102), (110), (103), (112) and (201) planes, respectively, and matched well with JCPDS card no. 75-1533 for the hexagonal wurtzite ZnO closely packed structure (*hcp* structure). A definite line broadening of the XRD peaks indicates that the prepared material consists of particles in the nanoscale range (Talam et al., 2012).

The crystallite size was calculated using the Debye Scherrer formula: $D = 0.9\lambda/\beta \cos \theta$, where D is the average crystallite size, λ is the wavelength of X-rays and β is the full width at half maximum for the most intense peak at 101 (Fig.

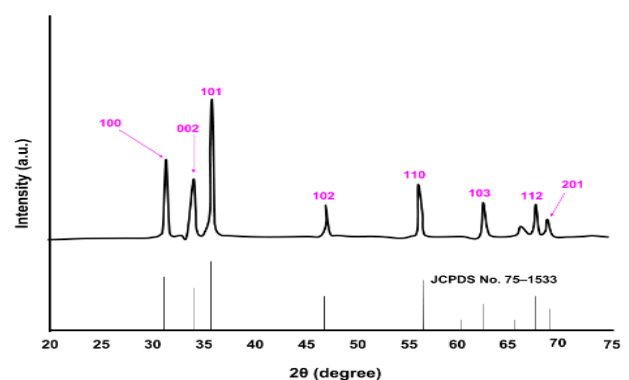


Fig. 1. X-ray diffraction (XRD) pattern of ZnO-NP.

1). The crystallite size was found to be ~9 nm. Crystallite size generally corresponds to the coherent volume in the material for the respective diffraction peak, it may also correspond to the grain size of a powdered sample, or a measure of the thickness of polycrystalline thin film or bulk material (Gubicza, 2012). The average size of ZnO-NP calculated was 8 nm which is in agreement with the acceptable range of 4–8 nm derived from the Scherrer equation and agrees with findings by Deepracha *et al.*, (2019). Hydrogen storage properties of nanomaterials were investigated by Andrew, (2019), the outcome shows the smaller the crystallite sizes, the reduced diffusion path of hydrogen enhance the sorption kinetics significantly.

The specific surface area of the synthesized ZnO-NP was examined using the Brunauer–Emmett–Teller (BET) method employing a BET analyzer as displayed in Figures 2 and 3 (a, b). Many of the unique, intrinsic properties associated with nanomaterials arise from the large surface-to-volume ratio of these exceptionally small materials. Surface area properties may also be related to environmental fate and hazard implications; therefore, accurately measuring surface area is extremely important for material characterization. The most commonly used method of measuring the surface area of nanomaterials is the Brunauer–Emmett–Teller (BET) surface adsorption method (Brame and Griggs, 2016).

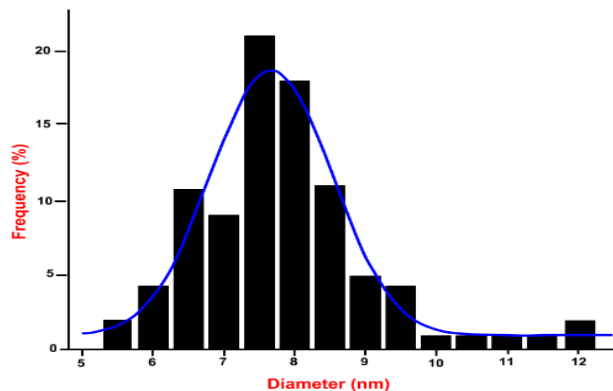


Fig. 2. Size distribution of ZnO-NP.

From Fig. 3, the pore size distributions indicated that ZnO presents a relatively narrow distribution ranging from 5 nm to 12 nm. Taking into account the morphology of the material revealed by the SEM micrograph, the small pores would represent the intra nanoparticles pores. The BET surface area of the synthesized ZnO was measured to be 39.8 m²/g. The pore size distribution curve was determined using Barrett, Joyner, and Halenda (BJH) methods as shown in Fig. 3 (b). The BJH average pore size, total pore volume and pore surface area of ZnO were recorded as 18.12 nm, 0.039 cm³/g and 11.67 m²/g, respectively. The sample exhibited isotherm of type IV (BDDT classification) with hysteresis loops of type H3 at relative pressure, indicating the presence of mesoporous structure, similarly as obtained

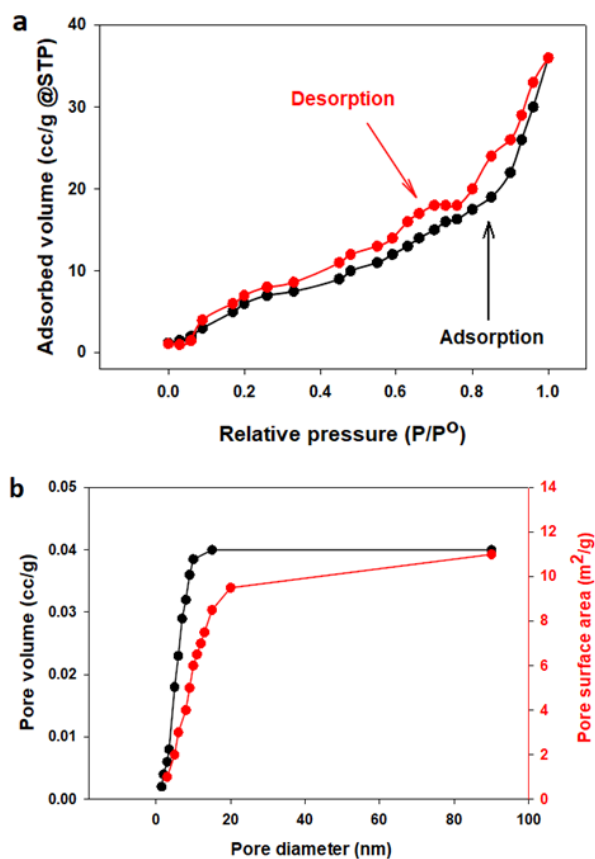


Fig. 3. Surface area (a) and Pore features (b) of the ZnO.

for TiO₂ by Wei *et al.*, (2013). The BET surface area of the prepared ZnO nanoparticles was 11.67 m²/g and the BET surface area of commercial P25 was 50 m²/g. A larger surface area provides more surface active sites for the adsorption of the reactive molecules, which leads the photocatalytic process to be more efficient. We can draw the conclusion that the nanoparticles prepared by us might have good photocatalytic activities.

Energy dispersion spectrum (EDS) reveals elemental weight compositions of 67.33% and 32.67% for Zn and O, respectively (Fig. 4), which are equivalent to 33.52:66.48 in atomic percentage. The prominence of Zn and O peaks in the spectrum confirm the formation of ZnO. The absence of

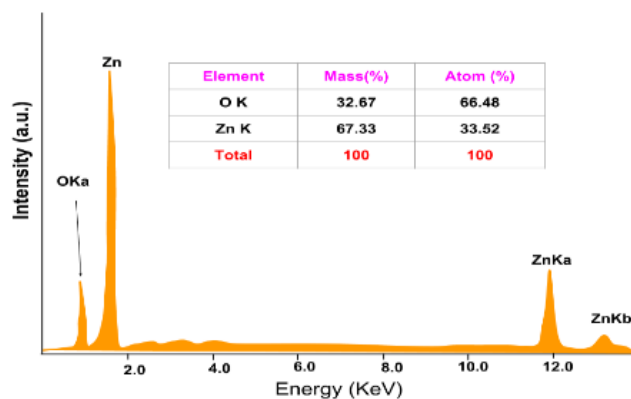


Fig. 4: EDS spectrum of ZnO-NPs.

any other elemental peak indicates the high purity of prepared samples. The nanoparticles appear to be spherical in shape with an average size of 10-130 nm. Recent studies have shown the application of chemical precipitation, adsorption, oxidation-reduction, evaporation, ionic exchange, electrochemical treatment, and membrane separation techniques in metal extraction from wastewaters (Krishna, 2018). Removal of toxic pollutants by adsorption on metal oxides has shown immense potential. Metal oxides possess large surface areas, porous structures, large number of active sites, thermal stability, easy recovery and low toxicity; they have shown outstanding performance for the adsorption and remediation of contaminants (Nagpal and Kakkar, 2018).

Textural morphology was confirmed by SEM micrograph as shown in Fig. 5. The micrograph shows that ZnO—NPs exhibited non-uniform agglomerated particles of lump-like spherical morphology. Owing to their size and higher surface area to volume ratio, nanomaterials possess some unique properties as compared to their bulk counterpart (e.g. charcoal), and the large surface area of these materials make them effective adsorbents in pollution remediation (Sarma et al., 2019). Fig. 3 further revealed well-developed mesopores (pores between 2 and 50 nm), and an easily modifiable surface, producing a large number of agglomerates of nanoparticles in nanocomposites (Rodríguez et al., 2020; Ashraf et al., 2018).

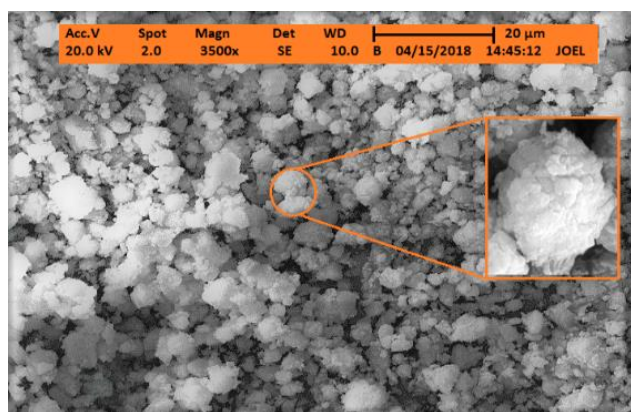


Fig. 5. SEM micrograph of ZnO-NP.

Fig. 6 is the graph of the photocatalytic behaviour of the adsorbent. Photocurrent response was recorded as a fast growth in which decay was visible. The ZnO-NP exhibited a stable photocurrent density of $3.35 \mu\text{A}/\text{cm}^2$; directly confirming that the adsorbent is capable of exhibiting high recombination efficiency of photogenerated electron-hole pairs. ZnO has been shown to offer multifarious advantages such as chemical-physical stability, low cost, and environmental friendliness (Deng et al., 2019). The photocurrent response is responsible for enhancing interfacial charge-transfer kinetics between the ZnO-NP and the adsorbate in the presence of wastewater, even though this may prove more effective in the removal of metals from

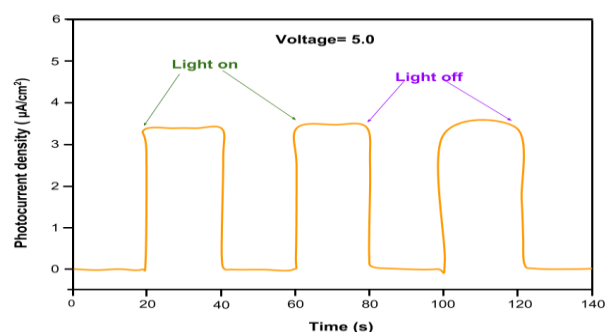


Fig. 6. Photocurrent response of ZnO-NP.

liquid wastes (Chandrasekharan and Kamat, 2000; Liu et al., 2018).

FTIR study was evaluated at wavelengths of $350\text{--}4000 \text{ cm}^{-1}$, absorption band appearing at region $3330\text{--}3600 \text{ cm}^{-1}$ is likely due to the O-H stretching of water molecule. The sharp bending vibration peak at 1522 and 1511 cm^{-1} was attributed to O-H stretching vibration of water. These vibrations represent the bound water on ZnO nanoparticles as displayed in Fig. 7. The sharp peak at 1486 cm^{-1} corresponded to the O-H stretching vibration (Mousavi et al., 2015). The absorption peaks at 968 , 602 and 439 cm^{-1} revealed the presence of Zn-O-H and Zn-O stretching vibration (Kansal et al., 2013).

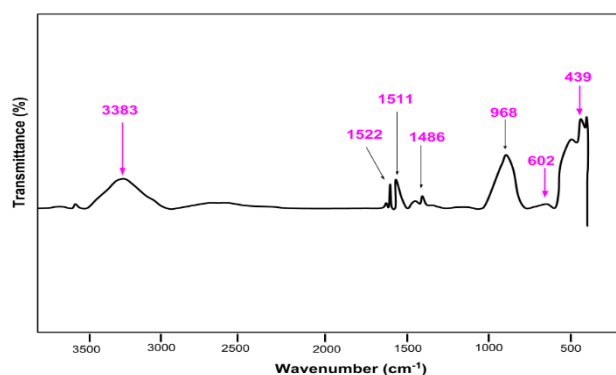


Fig. 7. FTIR spectra of ZnO-NP.

3.2 Comparative adsorptive removal of BOD and COD from NILEST tannery

The effect of adsorptive treatment time is presented in Tables 1 and 2. An initial concentration of 91.8 mg/L was documented for BOD at varying contact times of 20, 40, 60, 80 and 100 minutes. The BOD concentrations recorded after treatment for corresponding contact times were 17.0, 15.2, 13.0, 8.9 and 8.4 mg/L . The results recorded were below the WHO (world health organization) and NESREA (National Environmental Standards and Regulations Enforcement Agency) environmental limits of 30 mg/L . The percentage removal efficiency calculated and reported for the respective BOD concentrations was 81.5, 83.4, 85.8, 90.3 and 90.8 %. This is in close agreement with findings by Devi et al., (2008) on the reduction of chemical oxygen demand

(COD) and biological oxygen demand (BOD) of wastewater from coffee processing plant using activated carbon made from Avacado Peels with 99.02% and 99.35% removal at optimum operating conditions.

Table 1. Effect of treatment time on BOD and COD removal at pH 12 using 50.0 mg of ZnO-NP.

Contact time (min)	BOD Concentration (mg/L)	% Removal	COD Concentration (mg/L)	% Removal
20	17.0±0.10	81.5±0.01	502±0.08	12.8±0.01
40	15.20±0.10	83.4±0.02	480±0.02	16.7±0.01
60	13.0±0.01	85.8±0.02	416±0.11	27.8±0.02
80	8.90±0.08	90.3±0.30	352±0.01	38.9±0.08
100	8.40±0.01	90.8±0.11	288±0.01	50.0±0.01

The lowest percentage removal for COD was 12.8 %, and highest was 50 %. The removal of COD was mostly low (below 50 %) but improves progressively with increasing contact time. At 20 minutes intervals, COD reduction from an initial concentration of 576 mg/L decreased to 502, 480,

416, 352 and finally 288 mg/L; corresponding to contact times of 20, 40 60, 80 and 100 minutes respectively. Result reveals ZnO-NP has less removal preference for COD over BOD.

Results presented in Table 2 shows that increasing adsorbent dosage will influence BOD and COD removal efficiency at an optimum contact time of 30 minutes. The percentage removal of 79.2, 83.7, 88.6, 93.9 and 97.5 % on application of corresponding adsorbent masses of 20, 40, 60, 80 and 100 mg, for BOD at an initial concentration of 91.8 mg/L was studied (Table 2), with the result presenting the effect of adsorbent dosage on COD removal; percentage removal efficiency recorded ranged from a minimum of 27.8 to a maximum of 55.6 % at an initial concentration of 576 mg/L.

The decrease in BOD and COD with increasing adsorbent dose is ascribed to an increase in adsorption sites and surface areas of the ZnO-NPs, which led to an enhancement performance in the removal of the organic ions, due to the higher frequency of contact with the adsorbent, this is consistent with studies carried out by Singh *et al.*, (2013); Kumar, *et al.*, (2013), and Oladipo *et al.*, (2017). In a study, nanostructured ZnO semiconductor films were used for the degradation of organic contaminants (4-chlorocatechol) to a great effect.

Table 2. Effect of adsorbent dosage on BOD and COD removal at pH 12 and contact time of 30 min.

Mass of ZnO-NP (mg)	BOD treatment (mg/L)		% Removal	COD treatment (mg/L)		% Removal
	Before	After		Before	After	
20		19.1±0.01	79.2±0.11	416±0.01	27.8±0.09	
40		15.0±0.01	83.7±0.07	384±0.12	33.3±0.01	
60	91.8±0.10	10.5±0.05	88.6±0.11	576±0.01	320±0.09	44.4±0.01
80		5.6±0.01	93.9±0.01	288±0.01	50.0±0.06	
100		2.3±0.09	97.5±0.03	256±0.11	55.6±0.01	

4. Conclusion

This work has demonstrated the feasibility of synthesizing an eco-friendly adsorbent from aqueous extracts of the leaves of *Spinacia oleracea* and its application in reducing BOD and COD from tan liquor. Characteristics for the adsorptive potential of the adsorbent were evaluated and documented as having properties favourable for the removal of BOD and COD from tannery wastewaters. Application of ZnO-NP as an adsorbent for BOD and COD removal and its resultant efficiency shows great potentials while indicating the economic viability of diversifying the

usability of *Spinacia oleracea*. BOD removal was more effective over COD, therefore ZnO-NP is best efficient in the removal of BOD from tannery wastewater. Results also implicate the NILEST tannery wastewater as a large reservoir of biological and chemical oxygen depletant whose concentration does not satisfy the legal ranges for investigated parameters.

CRedit authorship contribution statement

D. P. Feka: Conceptualization, Methodology, Resources, Writing - original draft, Writing - review & editing. A. J. Habila: Methodology, Writing - review & editing. K. Francis:

Methodology, Data curation, Software. J. D. Habila: Writing - review & editing, Resources. M. Y. Bammai: Writing - review & editing

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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